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## Study on Water-Atmosphere Interface Distribution and Concentration Fluctuation Intensities of Polybrominated Diphenyl Ethers in Great Lake

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#### Abstract

In this study, the water-atmosphere interfaces distribution and concentration fluctuation intensities of polybrominated diphenyl ethers (PBDEs) in Great Lake were analyzed. Based on the selected significant quantum descriptors, bromination pattern descriptors ( $N_{PBS}$ ) and temperature descriptors by combination of variable important for projection (VIP) and correlation analysis, 35 experimental Henry's law constants of PBDEs (transformed as lgH, represented the water-atmosphere interfaces distribution) at different temperatures were divided into modeling set (25 samples) and testing set (10 samples) to establish the quantitative structureparameter relationship (OSPR) model for lgH at different temperatures via the optimal subset method firstly. Then, the Henry's law constants of other unknown PBDEs at different temperatures were predicted and the water-atmosphere interfaces distribution regularity were analyzed from views of chemical quantum descriptors and substituent pattern respectively. At last, the concentration fluctuation intensities in Great Lake were discussed based on the correlation analysis between the detected concentration data and experimental Henry's law constants. The established optimal QSPR model has shown a good fitness and predictive ability for Henry' law constants of PBDEs, with the  $R^2$  and  $R_{pred}^2$  of 0.985 and 0.978 respectively. From the view of chemical quantum descriptors, the lgH of PBDEs is positive correlative with  $E_{HOMO}$ , indicating the ability to provide electron of congener molecule has played a leading role in governing the exchange ability between water-atmosphere interface; from the view of bromination pattern, both number of relative positions and each position have great influence on the water-atmosphere interface distribution, especially  $N_{2(6)}$  and  $N_0$  can decrease the release of PBDEs from water to atmosphere in Great Lake. The concentration fluctuation intensity is negative correlated with the lgH linearly, the bigger solubility of congener in Great Lake, and the stronger water-atmosphere interface exchange ability and fluctuation intensity.

**Keywords:** polybrominated diphenyl ethers, quantitative structure-parameter relationship, Henry's law constants, water-atmosphere interface distribution, concentration fluctuation intensity

#### Introduction

Generally added to various industrial materials and home furnishings as additive brominated flame retardants, Polybrominated diphenyl ethers (PBDEs) are physically combinated with preproduced polymer simply and do not react with them chemically, so they are more likely to leach or volatilize from products and disperse in each medium (Labunska et al., 2013; Na et al., 2013).

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The widespread use of PBDEs increased their worldwide presence in various biotic and abiotic environment, even in the remote areas of the Arctic and Antarctica, producing significant negative effect on nervous system, endocrine system and immune system of wildlife and human (Mikula and Svobodova, 2006; Möller et al., 2011).

Most of the workdwide demand for PBDEs is distributed between North American (NA, 40%), the Far East (30%), and Europe (25) (Manchester-Neesvig et al., 2001). In Great Lake, Luckey et al (2001) has found that the concentration of total PBDEs was located in 4~13pg/L in NA which composing BDE-49 and BDE-99 mostly (>90%). Following research has also found that the large usage of the PentaBDE mixture can be seen in the ubiquitous presence of the congeners BDE-47, BDE-99 and BDE-100 in freshwater and marine environments in NA (Hites 2004; Zhu and Hites 2004; Yogui and Sericano, 2009). Both the penta- and octa-BDEs have been phased out of the NA markets while the Deca-BDE will have been completely withdrawn from the market by the end of 2013 (Kemmlein et al, 2009; Ma et al., 2013). The total PBDE concentrations in Great Lakes has presented increasing trend from 1989 to 2003 (Hites 2004; Zhu and Hites 2004), but current studies has shown an obvious decreased breakpoint that in concentration after 2000-2001 for Lake Huron, Michigan and Ontario according the statistical data (Crimmins et al, 2012). However, the concentration fluctuation intensities and corresponding regularity of PBDEs in Great Lake has never been analyzed in previous studies.

Physicochemical properties, especially those partitioning properties such as Henry' law constants (H), vapor pressure and  $K_{OA}$ , play a major role in their transport and mobility in the global environment. The Henry's law constants is an important indicator that plays a fundamental role in predicting the transport, behavior and fate of organic compounds in the environment which model the chemical transfer between atmosphere and water (Cetin et al., 2006). Meanwhile, Henry's law constants obtained from the Clapeyron equation are only valid for dilute concentrations (compounds with slightly or even moderately soluble in water), approximated by the ratio of the compound's vapor pressure and its aqueous solubility, but there are large deviations from practical solutions.

PBDEs contain 209 congeners distinguished by bromine number and substitution position (Palm et al., 2002; Gu et al., 2009). Compared with other POPs such as PCDD/Fs or PCBs, physicalchemical properties of PBDEs have been measured rather scare, meanwhile only fewer than 40 PBDEs congeners have been synthesized which makes it cumbersome and expensive to detect the physical and chemical properties of each compound (Jiang et al., 2012). In order to assess overall effect and variation rules of PBDEs in Great Lake more comprehensively, it is important to develop predictive models for physicochemical properties of PBDEs from their molecular structures. Quantitative structure-property relationship (QSPR) model provides an opportunity of studying a few compounds to predict properties of other unknown homolog (Blotevogel et al., 2011). Some previous studies showed that it was indeed feasible to predict the properties or activities with QSPR models for PBDEs (Niu et al., 2006; Chen et al., 2007; Papa et al., 2011), However, this kind of studies on Henry's law constants of PBDEs is still limited. Xu et al. (2007) and Papa et al. (2009) have used different descriptors to estimate the QSPR model for Henry's law constants of 7 PBDEs congeners, however, with less data, not take temperature into consideration, most selected descriptors reflects the overall performance of PBDEs molecular, no experimental data were used for validation and not combine with the actual sea area.

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Based on the existing problem, the aim in this study is to analyze the regularity of wateratmosphere interface distribution and concentration fluctuation intensities of PBDEs in Great Lake. We screen the decisive descriptors and analyze corresponding modularatory effect on water-atmosphere interface distribution of PBDEs in Great Lake firstly, and then establish the QSPR model for lgH of PBDEs at different temperatures to analyze the distribution regularity among 209 congeners. At last, the regularity of concentration fluctuation intensities in Great Lake will be discussed.

#### **Methods and Materials**

#### Experiment data source

Henry's law constants (transformed as lg*H* in this study for analysis conveniently) of seven PBDE congeners (BDE-28, BDE-47, BDE-99, BDE-100, BDE-153, BDE-154, BDE-209) were measured at five different temperatures during 5-40 °C using a gas-stripping technique by Cetin and Odabasi (2005), listing in Table 1 in supporting information. The concentration values for BDE-47, BDE-99, BDE-100, BDE-153, and BDE-154 in water of Great Lakes (Lake Erie, Lake Huron, Lake Michigan, Lake Ontario, Lake Superior) during 2004-2009 were detected by Crimmins (2012).

## Descriptors for QSPR modeling

A total of 22 descriptors were obtained to establish the QSPR model, containing quantum descriptors, bromination pattern descriptors (N<sub>PBS</sub>) and temperature descriptors. The molecular structures of all 209 congeners were initially optimized at the B3LYP/6-31(d) level by density function theory with Gaussian 09 software, then the frequency calculation was implemented at same level to detect whether optimized geometries were minimal on the potential energy surface (Worrall et al., 2004). Based the optimal geometries, the quantum descriptors were calculated and contained: the dipole moment( $\mu$ , debye), energy of the highest occupied molecular orbital( $E_{\rm HOMO}$ , eV), energy of the lowest unoccupied molecular orbital( $E_{\rm LUMO}$ , eV),  $E_{\rm LUMO}+E_{\rm HOMO}(E_{\rm S}, eV)$ , total energy (*TE*, eV), most negative atomic partial Mulliken charge in molecule( $q^{-}$ , e), most positive partial Mulliken charge in hydrogen atom (qH<sup>+</sup>, e), the mean polarizability ( $\alpha$ , 10<sup>-30</sup> esu) and the anisotropy polarisability ( $\Delta \alpha$ , 10<sup>-30</sup> esu).

Eight (N<sub>PBS</sub>) were presented as follow:  $N_{\rm T}$ , total bromine atom number;  $N_{2(6)}$ , bromine atom number at positions (2, 2', 6, 6');  $N_{3(5)}$ , bromine atom number at positions (3, 3', 5, 5'); N<sub>(4)</sub>, bromine atom number at positions (4, 4');  $N_{\rm o}$ , relative *ortho*-position of bromine atom;  $N_{\rm m}$ , relative *meta*-position of bromine atom;  $N_{\rm p}$ , relative *para*-position of bromine atom;  $N_{\rm D}$ , different number of bromine atom between two benzene rings. Five different forms of temperature descriptors (T,  $T^2$ , 1/T,  $\lg T$ ,  $T^3$ ) were used to establish the QSPR model.



Figure 1. Atom mark number and molecular geometry of PBDEs.

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## Method of QSPR modeling

35 known experimental samples of  $\lg H$  were divided into modeling set (25 samples) and testing set (10 samples) by interval-sampling method. Based on the modeling set, the correlation analysis between  $\lg H$  and 22 descriptors was implemented firstly to eliminate uncorrelated variables in the view of statistical point which didn't have significant effect on the wateratmosphere interface distribution; the correlation among independent variables had also been obtained coinstantaneously to provide guideline for eliminating collinearity in next step. Then, the variable importance for projection (VIP) analysis for remainder descriptors on  $\lg H$  was carried out by partial least square (PLS) regression to screen significant descriptors further whose VIP values were larger than 1 (Li et al., 2013).

Based on the screening significant descriptors, the best subset regression was used by Minitab software to establish a series of QSPR models. Among them, the optimal QSPR model was evaluated and decided by the comprehensive performance (fitness, robustness, predictability) of each model. For all established QSPR models, the conventional square of correlation coefficient ( $R^2$ ) was used to test the fitness of the models. The robustness was evaluated during leave-one-out (LOO) cross validation (Wold, 1978), which finally gave the cross-validatated correlation coefficient ( $q^2$ ) and the PRESS which was defined as the sum of squared differences between the experimental ( $y_i$ ) and predicted lgH ( $\hat{y}_i$ ) of the excluded observation in LOO treatment. The  $q^2$  and *PRESS* were calculated as follow:

$$q^{2} = 1 - \frac{PRESS}{SS} = 1 - \frac{\sum_{i=1}^{n} (y_{i} - \hat{y}_{i})^{2}}{\sum_{i=1}^{n} (y_{i} - \bar{y}_{i})^{2}}$$
(1)

Where *SS* stood for the residual sum of squares of the experimental values around the mean value ( $\bar{y}_i$ ). When  $q^2$  was larger than 0.5, the model was considered to have well robustness. The correlation coefficient produced by testing set ( $R_{pre}^2$ ) (Gramatica, 2007; Puzyn et al., 2008) was used to reflect the external predictive power of the model.

During the QSPR modeling, there was an annoying problem: more independence variables always gave rise to poorer accuracy; meanwhile too less always produced biased estimate. The parameter of Cp embedded in the best subset regression would provide help to select the best candidate regression model by weighing the balance. The smaller value of Cp always emerged the better performance of model.

## Concentration fluctuation intensities calculation

In order to evaluate the concentration fluctuation intensity of each congener in Great Lake conveniently, a fluctuation intensity index was put forward in this study. For all five congeners with known experimental concentration in Great Lake during 2004-2009, the calculation process was same and presented as follow:

$$y_{ij} = x_{ij} / x_{(i-1)j}$$
 (2)

Where  $x_{ij}$  and  $x_{(i-1)j}$  were the concentration of each congener in *i*th year and the last year (i=2005, 2006, 2007, 2008, 2009) in *j*th lake (j=1, 2, 3, 4, 5), the  $y_{ij}$  was the growth rate in *i*th year in *j*th lake.

$$\sigma_{j} = \sum y_{ij} / 5 \tag{3}$$

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Where the  $\sigma_i$  was the standard deviation (SD) of growth rate in *j*th lake during 2004-2009.

 $F=\sum \sigma_i / 5$ 

(4)

Where the F was the concentration fluctuation intensity index for each congener in Great Lake during 2004-2009.

## **Results and Discussions**

Descriptors screening and effect analysis for water-atmosphere interface distribution

Before QSPR modeling, the correlation analysis was performed for modeling set to illuminate general correlation between lg*H* and each descriptor (Table 1). It is found that most descriptors have strong correlation with lg*H* (sig.<0.01), but more serious collinearity between independent variables also exists at the same time (Table 2 in supporting information). Six descriptors ( $T^3$ ,  $\Delta E$ ,  $\mu$ ,  $q^-$ ,  $qH^+$ ,  $N_D$ ) with less importance (sig.>0.01) and constant variable  $N_4$  were excluded during following calculations.

Among the preliminary screening descriptors, the VIP values of all temperature descriptors,  $E_{\text{HOMO}}$ ,  $N_{2(6)}$ , and  $N_0$  are bigger than 1 (Table 1), representing significant effect on regulating the water-atmosphere interfaces distribution of PBDEs in Great Lake for these descriptors. Combined with the symbol of correlation analysis, the lg*H* is positive correlated with  $E_{\text{HOMO}}$  and negative with  $N_{2(6)}$  and  $N_0$  significantly.

Previous study has also found that increasing HOMO energy will result in higher reactivity (Niu et al., 2006; Fang et al, 2008). The greater the  $E_{HOMO}$ , the greater the ability of PBDE molecule to provide electrons, and the stronger the interactions between PBDE molecule and water molecule, thus the greater lg*H* and stronger ability of releasing from water to atmosphere, which is similar with the conclusion of Xu et al. (2007) and Papa et al. (2009).

Both the number of relative positions and substituent position has great influence on wateratmosphere interface distribution of PBDE congeners in Great Lake. The congeners with bromine atoms in *ortho*-position relative to the ether-link have lower vapor pressures compared to those with bromine atoms in the *meta/para*-positions, conductive to transport from water to gas, which may relative with the perpendicular molecular configuration, because only the congener configuration with *ortho*-position is perpendicular to each other for two benzene rings.  $N_0$  represents the pairs of bromine atoms in relative *ortho*-position in single benzene ring of molecule. The more  $N_0$ , the more concentrated distribution of bromine atoms in molecule and the more asymmetric charge distribution of molecule which is conductive to interact with water molecule and contain the release of congener to gas.

For other descriptors, the VIP values are all located in 0.90-0.1 except  $N_{\rm m}$ , all showing a certain degree effect on distribution between air and water interface. In order to avoid the mission of useful information and introduction of duplicate message, all the preliminary screening descriptors were filtrated ultimately by the conjoint comparison of correlation coefficient ( $R^2$ ) and VIP value, then seven descriptors: T,  $E_{\rm HOMO}$ ,  $\Delta E$ ,  $E_{\rm S}$ ,  $\alpha$ ,  $N_{2,6}$  and  $N_{\rm o}$  were screened for QSPR modeling at last.

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Table 1 Correlation analysis and VIP value of each descriptor of PBDEs.

Descriptors	Т	$T^2$	1/T	lgT	$T^3$	TE	$E_{\rm HOMO}$	$E_{\rm LUMO}$	$\Delta E$	$E_{\rm s}$	μ
$R^2$	0.612**	0.541**	- 0.626**	0.647**	0.472*	0.690**	0.737**	0.651**	0.471**	0.694**	0.372
sig.	0.001	0.005	0.001	0.000	0.017	0.000	0.000	0.000	0.017	0.000	0.067
VIP	1.18	1.12	1.11	1.16	-	0.93	1.14	0.90	0.94	0.94	-
Descriptors	$q^{-}$	$q{ m H}^{\scriptscriptstyle +}$	α	$N_{\mathrm{T}}$	$N_{2(6)}$	$N_{3(5)}$	$N_4$	$N_{ m o}$	$N_{ m m}$	$N_{\rm p}$	$N_{\rm D}$
$R^2$	-0.188	0.431*	- 0.685**	- 0.690**	- 0.675**	- 0.591**	-	- 0.587**	- 0.527**	- 0.644**	0.470*
sig.	0.368	0.032	0.000	0.000	0.000	0.002	-	0.002	0.007	0.001	0.018
VIP	-	-	0.92	0.93	1.09	0.92	-	1.06	0.82	0.90	-

\*\* Correlation is significant at the 0.01 level; \* Correlation is significant at the 0.05 level.

#### QSPR modeling and regularity analysis for water-atmosphere interface distribution

Based on the screening seven descriptors ultimately and modeling set, models (5-17) have been established by best subset regression, listing in Table 2. Through the comprehensive comparison for model performance, the model (11) has the best fitness, robustness, predictability and the smallest *C*p, being selected as the optimal QSPR model with the expression as follow:

$$lgH = 6.377 + 0.038T + 57.47E_{HOMO} - 0.021\alpha - 0.048N_{o}$$
(11)

As shown in Fig.2, for the training set and testing set, the points tends to cluster along the 45 ° tangent line indicating the predicted values are consistent with the corresponding determined values. Based on QSPR model (11), Henry' law constants for other 202 PBDEs at 5 different temperatures (Table 1 in supporting information) were predicted which may be useful for exposure assessment of these compounds in Great Lake. Our results suit very well for previous efforts. Cetin and Odabasi (2005) has reported the fitted regression equation for lg*H* with the average  $R^2$  0.958, which is smaller than our study (0.984).

No.	$R^2$	Ср	PRESS	$q^2$	$R_{\rm pred}^2$	Т	$E_{\rm HOMO}$	$\Delta E$	$E_{\rm S}$	α	$N_{2(6)}$	No
(5)	0.524	558.7	6.995	0.459	0.683							
(6)	0.460	636.9	8.232	0.364	0.579							
(7)	0.955	33.4	0.711	0.945	0.944		$\checkmark$					
(8)	0.887	112.4	1.659	0.872	0.912						$\checkmark$	
(9)	0.971	14.8	0.452	0.965	0.966							
(10)	0.970	16.3	0.456	0.965	0.960							
(11)	0.984	2.1	0.093	0.984	0.978							
(12)	0.982	4.4	0.297	0.977	0.980							
(13)	0.983	4.1	0.290	0.978	0.979		$\checkmark$					
(14)	0.983	4.1	0.289	0.984	0.979							
(15)	0.982	6	0.338	0.974	0.979							
(16)	0.982	6	0.320	0.975	0.978		$\checkmark$					
(17)	0.982	8	0.367	0.972	0.979							
(18)	0.984		0.245	0.981	0.978	$T^2$	$\checkmark$					
(19)	0.981		0.299	0.977	0.976	1/T	$\checkmark$					
(20)	0.983		0.276	0.978	0.977	lg <i>T</i>	$\checkmark$					

Table 2. Modeling fitting results.

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The VIP values for four forms of temperature descriptors  $(T, T^2, 1/T, \lg T)$  are all greater than 1. Meanwhile, the last three temperature descriptors were introduced in model (11) to replace the descriptor T one by one, obtaining the QSPR model (18-20). The well performances of these models manifest that the water-atmosphere interface distribution of PBDEs in Great Lake is correlated with temperature but not with the concrete descriptor form significantly.



**Figure 2.** Plots of experimented lg*H* values vs. those predicted by QSPR model (11).



**Figure 3.** The correlation of lg*H* and total bromine number of PBDE congener

Concentration fluctuation intensities of PBDEs in Great Lake

The experimental concentration values and calculated concentration fluctuation intensity index (*F*) of five congeners with known experimental concentration in Great Lake during 2004-2009 were listed in Fig.1 (supporting information) Fig. 4 separately, showing the increasing trend of *F* along with the total number of bromine atom. Although the concentration of BDE-47 is the biggest in Great Lake, but not demonstrates the strongest concentration fluctuation intensity, presenting the water-atmosphere interface transport is not correlated with the existing concentration significantly. Due to low water temperature relatively in Great Lake (lower than  $4 \C$  in summer of central water surface), we take  $5 \C$  as the actual water temperature in Great Lake in this study because of the known detected concentration in  $5 \C$ . The scatter diagram between lg*H* at  $5 \C$  and concentration fluctuation intensity index was shown in Fig.4, obtaining 3 different fitted lines. All the linear, exponential and polynomial fitting have shown well fitness with correlation efficiencies of 0.679, 0.721 and 0.902 separately, but the explicit correlation should be confirm further through more monitoring data.



Figure 4. Scatter diagram between  $\lg H$  at 5  $\degree$  and concentration fluctuation intensities of five PBDE congeners in Great Lake.

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## Conclusions

Based on the consideration of taking Henry's law constant as index of water-atmosphere interface distribution in Great Lake, the fitness, robustness and predictive ability of optimal QSPR model of lg*H* at different temperatures for PBDEs have all shown well performance, with the  $R^2$ ,  $q^2$  and  $R_{pred}^2$  of 0.985, 0.984 and 0.978 respectively. Among the selected descriptors, the water-atmosphere interface distribution of PBDEs in Great Lake are mainly regulated by temperature,  $E_{HOMO}$ ,  $N_{2(6)}$  and  $N_o$ , presenting the stronger ability to provide electrons of congener, perpendicular molecular configuration and asymmetric charge distribution are conductive to the releasing of PBDEs from water to atmosphere in Great Lake. Generally speaking, the concentration fluctuation intensities of five known detected congeners are negative correlated with the lg*H*, with correlation efficiencies of 0.679, 0.721 and 0.902 for linear, exponential and polynomial fitting. The bigger solubility of congener in Great Lake, the stronger water-atmosphere interface exchange ability and fluctuation intensity

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# References

- Blotevogel J., Mayeno A.N., Sale T.C. and Borch T. (2011). Prediction of contaminant persistence in aqueous phase: a quantum chemical approach. *Environmental Science & Technology*, 45(6), 2236-2242.
- Cetin B. and Odabasi M. (2005). Measurement of Henry's law constantss of seven polybrominated diphenyl ether (PBDE) congeners as a function of temperature. *Atomspheric Environment.*, 39(29), 5273-5280.
- Cetin B., Ozer S., Sofuoglu A. and Odabasi, M. (2006). Determination of Henry's law constantss of organochlorine pesticides in deionized and saline water as a function of temperature. *Atmospheric Environment.*, 40(24), 4538-4546.
- Chen J.W., Wang D.G., Wang S.L., Qiao X.L. and Huang L.P. (2007). Quantitative structure-property relationships for direct photolysis of polybrominated diphenyl ethers. *Ecxicology and Environmental Safety.*, 66(3), 348-352.
- Crimmins B.S., Pagano J.J., Xia X.Y., Hopke P.K., Milligan M.S. and Holsen T.M (2012). Polybrominated diphenyl ethers (PBDEs): turning the corner in Greak Lakes trout 1980-2009. *Environmental Science & Technology*, 46(18), 9890-9897.
- Fang L., Huang J., Yu G. and Wang L.N. (2008). Photochemical degradation of six polybrominated diphenyl ether congeners under ultraviolet irradiation in hexane. *Chemosphere*. 71(2), 258-267.
- Gramatica P. 2(007). Principles of QSAR models validation L internal and external. *QSAR & Combinatorial Science.*, 26(5), 694-701.
- Gu C.G., Ju X.H., Jiang X., Wang F., Yang S.G. and Sun, C. (2009). DFT study on the bromination pattern dependence of electronic properties and their validity in quantitative structure-activity relationships of polybrominated diphenyl ethers. *SAR QSAR Environmental Research*. 20(3-4), 287-307.
- Hites R.A. (2004). Polybrominated diphenyl ethers in the environment and people: A meta-analysis of concentrations. *Environmental Science & Technology*, 38(4), 945-956.
- Jiang Y.F., Wang X.T., Zhu K., Wu M.H., Sheng G.Y. and Fu J.M. (2012). Occurrence, compositional patterns, and possible sources of polybrominated diphenyl ethers in agricultural soil of Shanghai, China. *Chemoshpere*, 89(9), 936-943.
- Kemmlein S., Herzke D. and Law R.J. (2009). Brominated flame retardants in the European chemicals policy of REACH-Regulation and determination in materials. *J. Chromatogr, A*, 1216(3), 320-333.

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- Labunska I., Harrad S., Santillo D., Johnston P. and Brigden K. (2013). Levels and distribution of polybrominated diphenyl ethers in soil, sediment and dust samples collected from various electronic waste recycling sites with Guiyu town, southern China. *Environmental Science: Processes & Impacts*, 15(2), 503-511.
- Li Y., Jiang L., Li X.L., Hu Y. and Wen J.Y. (2013). A QICAR model for metal ion toxicity established via PLS method. *Chem.Res. Chin. Univ.*, 29(3), 568-573.
- Luckey F., Fowler B. and Litten S. (2001). Establishing baseline levels of polybrominated diphenyl ethers in Lake Ontario surface waters In: Proceedings of the second international workshop on brominated flame retardants. The Swedish chemical society.
- Ma Y.N., Salamova A., Venier M. and Hites R.A. (2013). Has the phase-out of PBDEs affected their atmospheric levels? Trends of PBDEs and their replacements in the Great Lakes atmosphere. *Environmental Science & Technology*, 47(20), 11457-11464.
- Manchester-Neesvig J.B., Valters K. and Sonzogni W.C. (2001). Comparison of Polybrominated diphenyl ethers (PBDEs) and polychlorinated biphenyls (PCBs) in Lake Michigan Salmonids. *Environmental Science & Technology*, 35(6), 1072-1077.
- Mikula P. and Svobodova, Z. (2006). Brominated flame retardants in the environment: their sources and effects (a review). *Acta Veterinaria. Brno*, 75(4), 587–599.
- Möller A., Xie Z.Y., Sturm R., and Ebinghaus R. (2011). Polybrominated diphenyl ethers (PBDEs) and alternative vrominated flame retardants in air and seawater of the European Arctic. *Environmental Pollution.*, 159(6), 1577-1583.
- Na S., Kim M., Paek O. and Kim Y. (2013). Dietary assessment of human exposure to PBDEs in South Korea. *Chemosphere*. 90(5), 1736-1741.
- Niu J., Shen Z., Yang Z., Long X. and Yu G. (2006). Quantitative structure-property relationships on photodegradation of polybrominated diphenyl ethers. *Chemosphere*. 64(4), 658-665.
- Palm I.T., Cousins D., Mackay M., Tysklind C. and Metacalfe M.A. (2002). Assessing the environmental fate of chemicals of emerging concern: a case of the polybrominated diphenyl ethers. *Environmental Pollution*, 117(2), 195–213.
- Papa E., Kovarich S. and Gramatica P. (2009). Development, Validation and inspection of the applicability domain of QSPR models for physicochemical properties of polybrominated diphenyl ethers. QSAR & Combinational Science. 28(8), 790-796.
- Papa E., Kovarich S. and Gramatica P. (2011). On the use of local and global QSPRs for the prediction of physicochemical properties of polybrominated diphenyl ethers. *Molecular Informatics*, 30(2-3), 232-240.
- Puzyn T., Suzuki N. and Haranczyk M., (2008). How do the partitioning properties of polyhalogenated POPs change when chlorine is replaced with bromine? *Environmental Science & Technology*, 42(14), 5189-5195.
- Wang Z.Y., Zeng X.L. and Zhai Z.C. (2008). Prediction of supercooled liquid vapor pressures and n-octanol/air partition constants for polybrominated diphenyl ethers by means of molecular descriptors from DFT method. *Science of the Total Environment*, 389(2-3), 296-305.
- Worrall F. and Thomsen M. (2004). Quantum vs. topological descriptors in the development of molecular models of groundwater pollution by pesticides. *Chemosphere.*, 54(4), 585-596.
- Wold S. (1978). Cross-validation estimation of the number of components in factor and principal components analysis. *Technometrics*, 20(4), 397-405.
- Xu H.Y., Zou J.W., Yu Q.S., Wang Y.H., Zhang J.Y. and Jin H.X. (2007). QSPA/QSAR models for prediction of the physicochemical properties and biological activity of polybrominated diphenyl ethers. *Chemosphere*, 66(10), 1998-2010.
- Yogui G.T. and Sericano J.L. (2009). Polybrominated diphenyl ether flame retardant in the US marine environment: A review. *Environment*al International, 35(3), 655-666.
- Zhu L.Y. and Hites R.A. 2004. Temporal trends and spatial distributions of brominated flame retardants in archived fishes from the Great Lakes. *Environmental Science & Technology*, 38(10), 2779-2784.